



Ag-TiO₂ Assisted Photocatalytic Degradation of Cytostatic Drug Cyclophosphamide Under UV-VIS Light

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Abstract: Cyclophosphamide (CP) is a commonly prescribed cytostatic drug that has cytotoxic, genotoxic, mutagenic, carcinogenic and teratogenic effects on living organisms, which asks for its elimination from any water source. The photocatalytic degradation of CP under UV-VIS irradiation was studied, using non-doped TiO₂ and Ag-doped TiO₂ (0.5-1.7% wt.). The catalysts were synthesized by vacuum deposition method. The influence of Ag concentration on the degradation performance of CP was investigated and an optimal content of 1.0% Ag was established. Using of this catalyst assures the pollutant degradation with a rate constant $k_{CP} = 6.59 \times 10^{-4} \text{ s}^{-1}$ and an efficiency $\eta_{CP} = 99\%$ after 120 min irradiation time. Based on the results of the quenching experiments in the presence of suitable scavengers, a mechanism of photocatalytic degradation of CP has been proposed. This consists in the attack of free hydroxyl radicals and superoxide radicals on the pollutant. Also, the contribution of Ag to inhibition of charge recombination and the additional generation of superoxide radicals, which are responsible for the higher photocatalytic activity of Ag doped TiO₂ compared to non-doped TiO₂, was emphasised.

Keywords: Cyclophosphamide, un-doped TiO₂, Ag-doped TiO₂, photocatalysis, reactive species

1. Introduction

Cytostatic drugs used in chemotherapy represent emerging pollutants which can affect quality of surface water that are receiving urban wastewater treatment plants (WWTPs) effluents. These pollutants reach the sewage through hospital effluents and domestic wastewater discharged.

Due to their specific characteristics (polar character, relative high water solubility and low absorption of sunlight with $\lambda > 290\text{nm}$) many of these pollutants are poorly photolytic converted under sunlight, weak biodegradable and not fully metabolised. These are the main reasons why cytostatic drugs resist to classical physical-chemical and biological processes applied within WWTPs [1].

The cytostatic drugs concentration in water bodies depends on their type and consumption level. In the case of cyclophosphamide (CP), compound belonging to oxazaphosphorines class which is one of the most commonly prescribed antineoplastic drug, the reported concentration in water and water bodies is usually under 30 ng/L in EU countries [2-6], excepting Romania which identified CP concentration twice higher in Somes river [7]. Anyway, the concentration level is in good correlation with statistic data regarding CP consumption, which in UE member states is around hundreds kg/years/country [8].

Since CP, like other cytostatic drugs, is intended to disrupt or prevent cellular growth, it presents potential cytotoxic, genotoxic, mutagenic, carcinogenic, and teratogenic effects on living organisms. For these reasons, cytostatic drugs should be removed from water.

Application of advanced oxidation processes (AOPs) may be a possible solution of this problem. There are examples about antineoplastics removal from various water sources by ozonation [9], H₂O₂

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assisted photolysis, photo-Fenton process [10] and photocatalysis assisted by commercial TiO₂ [11-13]. The reported data reveals that efficient pollutant degradation occurs in optimum working conditions (pH, H₂O₂ and Fe²⁺ concentrations, TiO₂ dose, cytostatic drug concentration, and irradiation time).

Although, TiO₂ photocatalyst presents many advantages like: enhanced chemical stability, low toxicity, and relatively low cost [14], it is photoactive only under UV irradiation, due to its large band gap energy (3.2eV for anatase). Another drawback of pristine *titania* is related to fast recombination of photo-generated electrons and holes which has as result the decrease of oxidizing reactive species concentration with negative impact upon pollutant degradation efficiency. In order to solve these problems a lot of scientific work has been reported within the following research directions:

- shifting the TiO₂ photoactivity to the visible(VIS) range by doping of semiconductors with non-metals (N and S) [15-17];
- inhibition of charges recombination by doping of semiconductors with heavy metals like Fe, Ni, Co, Cr, V [18-22].

Besides these strategies, addition of noble metals like Pt and Ag to pristineTiO₂ [23, 24] has been proved to be an effective method to improve photocatalytic activity due to fact that noble metals act as traps for photo-generated electrons(e⁻) and inhibit their recombination with holes(h⁺). On the other hand, noble metals nano-particles present strong absorption in visible range and can improve sunlight photo-excitation of doped *titania* which assures the decrease of energy consumption and cost reduction.

Among noble metals doped *titania* only Pt-TiO₂ has been the subject of some systematic studies on CP photocatalytic degradation [25]. The results showed that Pt-TiO₂ assures a 5 times increase of rate constant, compared withpristineTiO₂.

Although, photocatalytic activity of Ag doped *titania* is less enhanced compared with Pt doped catalyst, Ag is most suitable for industrial application due to its low cost and easy preparation [26]. In addition, Ag has special behaviour related to oxygen adsorption and photocatalytic pollutant degradation is an oxidation processes controlled by the charge transfer to O₂.

In this paper we present results concerning Ag – doped TiO₂ assisted photocatalytic degradation of CP under UV-VIS light, a topic still not approached according to our knowledge. The photocatalytic activity of Ag-doped TiO₂ was compared with un - doped TiO₂ and the influence of Ag content on CP photocatalytic degradation efficiency and kinetics were studied. All catalysts were synthesised by vacuum deposition method. The identification of reactive species involved in CP photocatalytic conversion was also investigated by quenching experiments performed in the presence of hydroxyl radicals, superoxide radicals and holes scavengers, in order to elucidate the mechanism of pollutant degradation on Ag - doped TiO₂.

2. Experimental part

2.1. Materials

The following reagents were used: Cyclophosphamide (Alfa Aesar) 97% purity; Trifluoroacetic anhydride (Sigma Aldrich) 99% purity, Toluene (Merck) GC-MS grade, Sodium chloride (Chempur) 99,5% purity, Sodium sulphate anhydrous (Chempur) 99% purity.

The following reagents were used as scavengers for reactive species: Isopropanol (J.T.Baker) ACS grade, Sodium fluoride (Chempur) 99 % purity, Ethylenediaminetetraacetate disodium salt (Merck) 98.5% purity, Benzoquinone (Sigma Aldrich) 98% purity.

All aqueous solutions were prepared using ultrapure water generated by a MilliQ Integral 15 equipment.

Un-doped TiO₂ and Ag-doped TiO₂ with various metal concentrations (0.5-1.7wt%) catalysts (synthesised by SC MGM STAR CONSTRUCT SRL) were used. Each of them has been deposited as thin film (300nm) on glass plate with 14 cm² area.



2.2. Experimental set-up

The photodegradation experiments were performed using a batch reactor with medium-pressure mercury lamp TQ-150-Z3 which emits radiation in the domain $\lambda = 320\text{-}500\text{ nm}$. The lamp equipped with an glass water cooling jacket and 6 plates of catalyst placed equidistantly around the light source were immersed in the centre of the reactor ($V = 1000\text{ mL}$) containing the CP solution. The determination of photon's flow of emitted radiations was performed by chemical actinometry using ferrioxalate solution and a value of $I_0 = 1.05 \times 10^{-6}\text{ einstein s}^{-1}$ was registered. Using various catalysts, solution with 0.04 mM CP content was photo-oxidized in the following working conditions: $\text{pH} = 8.5$; catalyst total irradiation area = 84 cm^2 ; irradiation time = $30 - 120$ minutes. Prior to irradiation (30 min.) and during photocatalysis experiments, air was bubbled into solution at 50 L/h flow.

All experiments were duplicated in order to assure data quality.

2.3. Analytical method

In order to assess the efficiencies of pollutant degradation, initial and irradiated samples were analysed for CP concentration by gas chromatography coupled with ion trap mass spectrometry (GC-ITMS) using an Agilent 7890A – Agilent 240MS equipment. Volumes of 10 mL of samples were saturated with sodium chloride. The samples pH was adjusted to 10 by addition of 0.5 mL of 1 M ammonia buffer solution. The treated samples were extracted with $2\text{-}5\text{ mL}$ of toluene. Organic extract was dried with anhydrous Na_2SO_4 . 1 mL of extract was placed in 2 mL GC vial and then $10\text{ }\mu\text{L}$ of trifluoroacetic anhydride was added. After sealing and shaking the vials were heated at 70°C for 2 h . After cooling, $1\text{ }\mu\text{L}$ of treated samples were injected and analysed by GC-ITMS. For separation, low polarity phase column (5% diphenyl/ 95% dimethyl polysiloxane, 30 m , 0.25 mm ID, $0.25\text{ }\mu\text{m}$ film) was used. Retention time for CP derivative was 11.02 min , GC conditions were: split injection ($1:10$) with injector temperature 280°C , flow rate: 2 mL/min ; carrier gas: helium 6.0 ; temperature ramp employed: 50°C (1 min), 20°C/min to 280°C , hold 3.5 min . Mass detector conditions: Ion Trap: 120°C ; Manifold: 50°C ; Transfer line: 280°C ; Ion source: 250°C . Scan mode: single ion monitoring (quantified ion: $m/z = 307$ and qualifier ion: $m/z = 309$).

3. Results and discussions

3.1. Photocatalytic activity

The capture of photo-generated electrons can retard electron/hole recombination during irradiation, thus exceeding the lifetime of charge carriers. Doping of TiO_2 with heavy or noble metal ions by scattering of metal nano-particles in titania matrix, that acts as electron trap, also assures improvement of photocatalytic activity [27]. The photocatalytic activity of un-doped TiO_2 and Ag - doped TiO_2 with various noble metal content, was investigated for CP degradation from aqueous solution (0.04 mM). The dopant concentrations, expressed as ratios related to TiO_2 were 0.5 , 1.0 and $1.7\text{ wt.}\%$. All experiments have been performed in aerobic condition at $\text{pH} = 8.5$, for 84 cm^2 catalyst total irradiation area and irradiation time between $30\text{-}120\text{ min}$. The efficiency of pollutant degradation was evaluated by determination of CP concentration vs. irradiation time in treated samples. Kinetic curves recorded for various Ag content compared with un-doped *titania*, linearized by a first order kinetic, are presented in Figure 1.

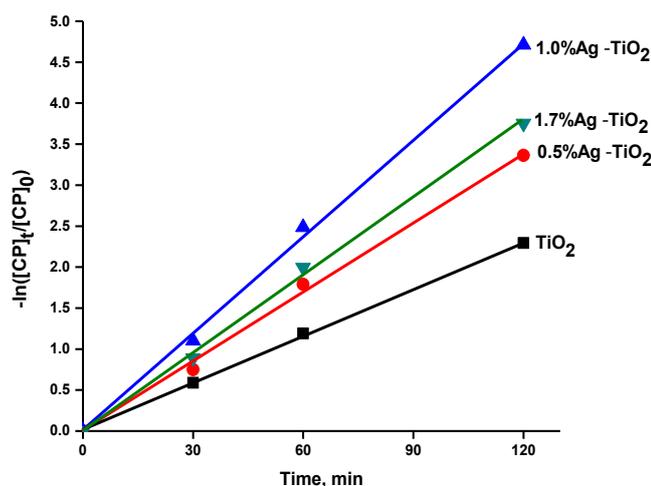


Figure 1. Linearized CP degradation kinetic curves of 0.5-1.7% Ag-doped TiO₂ compared with un-doped TiO₂

From the slope of each equation, pseudo-first order rate constants, k_{CP} were calculated. Pollutant degradation efficiencies resulted at 120 min. irradiation time and k_{CP} values for un-doped and Ag-doped TiO₂ with various metal content are presented in Table 1.

Table 1. CP degradation rate constants and efficiencies (120 min) on un-doped and Ag- doped TiO₂ with various dopant content

Catalyst type	[Ag] (wt.%)	$k_{CP} \times 10^4$ (s ⁻¹)	η_{CP} (%)
Un-doped TiO ₂	-	3.212±0.029	89.89
	0.5	4.703±0.098	96.53
Ag-doped TiO ₂	1.0	6.590±0.102	99.07
	1.7	5.263±0.084	97.66

Dopant presence within the catalyst matrix assures improvement of TiO₂ photocatalytic activity, the values of pollutant degradation rate constants being higher than those obtained for un-doped *titania*. As experimental results revealed, a relative low Ag content of 1.0%wt. allows the CP degradation rate constant increase twice, compared with un-doped TiO₂. However, exceeding of this dopant level, considered as optimum, has an opposite effect. Thus, the CP degradation rate constants decreased with 20%, if Ag concentration is 70% higher than optimum dopant level.

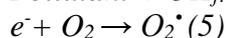
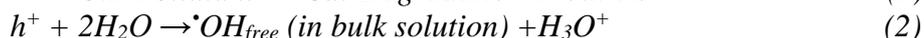
This behaviour is associated to the change of e^-/h^+ recombination rate which is influenced by dopant content. It is known that under irradiation, TiO₂ is activated and photo-generates e^-/h^+ pairs. When Ag is dispersed in semiconductor matrix, electrons in the conduction band of TiO₂ can be trapped by Ag nano-particles and inhibit the electrons –holes recombination, because the Fermi level of TiO₂ is higher than that of Ag clusters which became more reductive and behave as accumulation sites of photo-generated electrons transferred from *titania*. More Ag loaded in TiO₂ matrix increases the number of Ag cluster for electron trapping, which has as result a better charges separation and thus a better photocatalytic activity, as the experimental results showed. However, as more electrons are trapped on Ag clusters, increases the probability that these negative charged sites to become centre for positive charged holes capture. This behaviour affects the electron – holes separation, leading to the decrease of photocatalytic activity. In addition, higher Ag loaded would cover more TiO₂ surface and hinder the contact of catalyst with photons and pollutants, which also results in CP degradation rate constant decreasing, as is showed in the case of 1.7%Ag-doped TiO₂. Based on the experimental



results assessment it was established that 1% wt. Ag dopant content assures suitable CP photocatalytic degradation ($\eta_{CP} \geq 99\%$) after 120 min. irradiation time.

3.2. Effect of scavenger's presence and degradation mechanism

In the photocatalytic reactions assisted by catalyst, the main pathways of pollutant degradation are oxidation by direct holes (h^+) (Eq.1) and by free and adsorbed hydroxyl radicals generated from interaction between holes and H_2O ($\cdot OH_{free}$, in bulk solution) or $\cdot OH_{ads}$ ions adsorbed on catalyst surface (Eq. 2-4). Another possible oxidation route consists in reaction of pollutant with superoxide radicals ($O_2^{\cdot -}$) generated from electrons (e^-) trapping by molecular oxygen (Eqs.5- 6) [28].



In order to identify any possible contribution of the mentioned reactive species to CP photocatalytic degradation on Ag-doped TiO_2 , the influence of suitable scavengers on the cytostatic drug oxidation rate constant was investigated. Quenching experiments were performed on CP solution (0.04mM) in the presence of following scavengers:

- Disodium ethylenediaminetetracetate (EDTA, 4mM) for trapping of holes;
- Isopropanol (IPR, 4mM) for trapping of free hydroxyl radicals;
- Sodium fluoride (NaF, 2mM) for trapping of hydroxyl radicals adsorbed on catalyst surface;
- 1,4-benzoquinone (BQ, 4mM) for trapping of superoxide radicals.

The results of quenching experiments performed on un-doped TiO_2 and Ag-doped TiO_2 are presented in Table 2 and Figures 2, 3.

Table 2. CP degradation rate constants and efficiencies (120 min) on un-doped and Ag-doped TiO_2 in the presence of various reactive species scavengers

Catalyst type	Scavengers type	$k_{CP} \times 10^4 \text{ (s}^{-1}\text{)}$	$\eta_{CP} \text{ (%)}$
Un-doped TiO_2	-	3.212±0.029	89.89
	EDTA	2.933±0.055	87.60
	IPR	0.845±0.018	44.90
	NaF	2.792±0.072	86.20
	BQ	2.028±0.035	70.03
Ag-doped TiO_2	-	6.590±0.102	99.07
	EDTA	6.140±0.070	98.77
	IPR	2.650±0.063	84.56
	NaF	6.453±0.083	99.09
	BQ	3.330±0.076	90.60

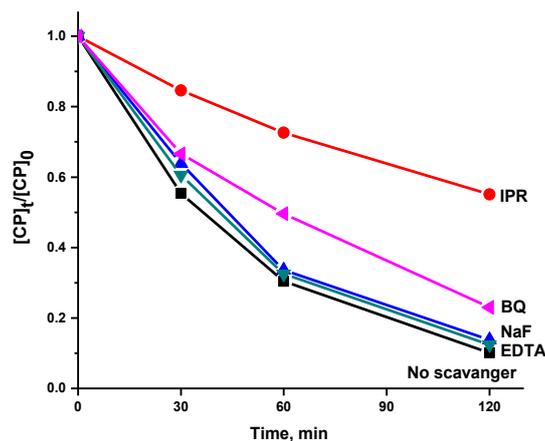


Figure 2. CP degradation kinetic curves on un-doped TiO₂ under UV-VIS irradiation with and without addition of various reactive species scavengers

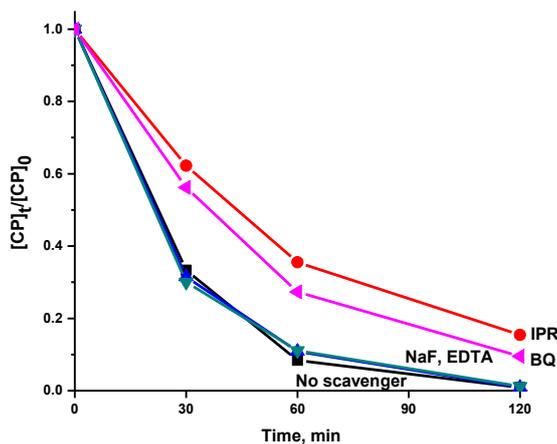


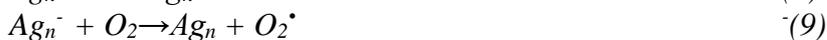
Figure 3. CP degradation kinetic curves on 1.0% Ag-doped TiO₂ under UV-VIS irradiation with and without addition of various reactive species scavengers

It is known that scavenger presence has as result the decrease of pollutant degradation rate constant due to reactive species quenching. This effect is different for various reactive species, thus the rate constant decrease in the presence of scavenger is more accentuated when a certain reactive species are more involved in the degradation process. In the case of un-doped TiO₂, the CP degradation takes place mainly by free hydroxide radicals, but superoxide radicals has also a contribution. Values of CP degradation rate constants in the presence of IPR and BQ showed the most pronounced decreases compared to the rate constant in the absence of scavengers, namely 74% (in the presence of IPR) and 37% (in the presence of BQ) respectively. Although, the reactive species involved in the degradation of the pollutant in the case of the Ag-doped TiO₂ are similar to those identified for the un-doped TiO₂, the contribution of superoxide radicals is higher in the case of noble metal-doped TiO₂. This behavior is due to the fact that Ag nano-particles are trapping sites for photo-generated electrons and concentration of generated superoxide radicals increases compared with un-doped TiO₂.

Based on the quenching experiments results and Ag content influence on CP degradation efficiency corroborated with literature information it was proposed a mechanism of reactive species generation



and their contribution to pollutant degradation by photocatalysis assisted by Ag-doped TiO₂ under UV-VIS illumination, which is described by following equations:



Under UV-VIS illumination of Ag-doped TiO₂, electrons are promoted to the conduction band (CB) and holes are generated in the valence band (VB) of TiO₂ (Eq.7). The photo-generated electrons are transferred to surface Ag cluster, which behave like accumulation sites (Eq.8). As experimental results showed, better charges separation is assured by Ag concentration increases up to optimum content. The trapped electrons react even with adsorbed molecular oxygen or with titanium surface ions (Ti⁴⁺) forming superoxide (O₂^{·-}) reactive species and reactive surface center (Ti³⁺), respectively, as is illustrated in Eqs. (9) and (10). Data reported in the literature showed that these reactions of trapped electrons slow down charge recombination and accelerate generation of superoxide ions and surface reactive centers [24]. On the other hand, the hydroxyl radicals level also increased, due to the increase of holes concentration implicated in their generation (Eq.11), as a consequence of electrons capture on Ag clusters. The main reactive species involved in CP degradation are free hydroxyl radicals (in bulk solution) and superoxide radicals (on catalyst surface) (Eq.12), as quenching experimental results showed. The possible oxidative conversion route is nucleophilic attack on chloroalkyl-side chain of pollutant as was reported in the case Pt or Fe-doped TiO₂ catalysts applied in the photocatalytic degradation of Ifosfamide, a structure isomer of CP [25, 29-31]. When Ag level overcomes the optimum content, the number of trapped electrons is high enough for Ag clusters to become recombination centers of holes. (Eq.13) and the CP degradation performance is detrimental affected, as the experimental results regarding the influence of Ag concentration on degradation efficiency showed.

4. Conclusions

CP degradation was studied using UV-VIS/Ag-doped TiO₂ systems. Experimental results showed that Ag loading increase of up to 1.0% wt. content had a positive influence on the pollutant degradation efficiency (η_{CP} = 99 %). Enhanced photocatalytic activity of Ag-doped TiO₂ compared with un-doped TiO₂ is a consequence of charges recombination inhibition, due to fact that nano-particles of Ag on TiO₂ surface acts as electrons captors. As quenching experiments revealed, the main reactive species involved in CP photocatalytic degradation were free hydroxyl radicals and superoxide radicals for both catalyst, but O₂^{·-} participation is more pronounced in the case of Ag-doped TiO₂ compared with un-doped TiO₂. It can be concluded that these reactive species together with reactive centre (Ti³⁺) formed on the catalyst surface, as the proposed mechanism showed, are both responsible for the enhanced photocatalytic activity of Ag-doped TiO₂. Experimental results proved that photocatalytic degradation under UV-VIS irradiation using Ag-doped TiO₂ is an attractive alternative for cytostatic drugs like CP.

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